

Comparison of Geospatial-Temporal Modeling Approaches in Air Pollution Estimations

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Abstract: Recent advancements in the geographic information systems and remote sensing technology have supported the development of geospatial-temporal modeling approaches for air pollution. Particulate matter (PM₁₀) and ozone (O₃) are two pollutants of great concern in all pollutants. Previous studies estimated the spatial-temporal variability of PM₁₀ and O₃ using a single model, but only a few studies considered exposure assessment using multiple models and compared model performance. In this study, PM₁₀ and O₃ data during 2015 to 2018 were collected from specific industrial monitoring stations provided by the Taiwan Environmental Protection Agency. Three geospatial-temporal modeling approaches including land-use regression (LUR), geographically weighted regression (GWR), and geographically and temporally weighted regression (GTWR) were used to predict PM₁₀ and O₃ exposure. Furthermore, the kriging-based hybrid model was integrated with these three geospatial-temporal models, and totally performs six models for each pollutant for our comparison. The results showed that integrating the GTWR and kriging-based hybrid models have the greatest performances compared to LUR, GWR, and the combination of both with kriging-based hybrid models. R² obtained from the GTWR coupled with kriging-based hybrid models for PM₁₀ and O₃ was 0.96 and 0.92, respectively. Of all variables used, wind speed, pure residential area, manufacturing, park; rice field, orchard; and forest land were important predictors for PM₁₀. Whereas, wind direction, industrial area, dry farming, and orchard were variables selected to predict O₃.

1. BACKGROUND AND AIM

Particulate matter with an aerodynamic diameter between 10-2.5 μm (PM_{10}) and ozone (O_3) exposure has been identified as a significant risk factor for the development of lung cancer and adverse health outcomes from cardiovascular and respiratory causes (Brook et al. 2010; Pope and Dockery 2006; Aguilera et al. 2015; Pope et al. 2002; Sabaliauskas et al. 2015). As personal monitoring is not generally feasible for large cohorts, methods to assess accurately within-city variability in exposure to PM_{10} and O_3 are required (Jerrett et al. 2005; Wu et al. 2017). In the past period, there are many predictive methods for capturing ambient air pollution gradients. Spatial interpolation, such as Kriging interpolation (Bayraktar and Turalioglu 2005), predicted pollutant level in an area from a limited number of monitoring sites. Spatial autocorrelation, the statistical relationships of distance among the measured points were used to explain and predict the variation of air pollutants in the surface. However, intra-urban air pollution concentrations could vary due to proximity to industrial parks, road and traffic density, and other site characteristics, such as population and land use (Tunno et al. 2016). The lack of consideration about the local emission sources between monitoring sites could deteriorate the accuracy of predictions. Compared with spatial interpolation, land-use regression (LUR) has been proved to have more advantages on characterizing the spatial relationships between local emissions and intra-urban pollution variations (Clougherty et al. 2013; Hoek et al. 2008; Michanowicz et al. 2016). LUR normally combines distributed pollution measures at multiple sites with a set of potentially predictive geographic source covariates, to develop a multiple linear regression model that can be rendered in a Geographic Information System (GIS) to estimate air pollution levels at unmeasured areas (Wu et al. 2017). Geographic predictors include traffic patterns, surrounding land-use allocations, demographic characteristics, green space distribution, and micro climatic conditions (Aguilera et al. 2015; Su et al. 2010; Wang et al. 2013; Wu et al. 2017; Shi et al. 2016). Likewise, geographically weighted regression (GWR) and geographically and temporally weighted regression (GTWR) are also spatial predictive methods for modelling spatial-temporal variation of air pollution, and gain more and more attention in recent studies (Chu et al., 2018; Cui et al., 2019; Guo et al., 2017; Ma et al., 2018).

In this study, three geospatial-temporal modeling approaches including LUR, GWR), and GTWR were used to predict PM_{10} and O_3 exposure. Furthermore, the kriging-based hybrid model was integrated with these three geospatial-temporal models, and totally performs six models were performed for each pollutant for our comparison.

2. METHODS

The coastal areas of Kaohsiung City, in which several heavy metal industrial parks located on, were selected for the experimental study (Fig. 1). Monthly averaged concentrations from May 2015 to September 2018 of the two study pollutants, PM_{10} and O_3 , were obtained from 9 specific industrial monitoring stations. There are four steps for data Analysis (Fig. 2). Environmental factors such as meteorological factors (relative humidity, and temperature etc.),

co-pollutants (SO₂, and PM_{2.5} etc.), topography (elevation), land-use distributions (residential areas, and surrounding greenness etc.), and community emission sources (temple and Chinese restaurant) were combined with the recorded concentrations and used as explanatory predictors to develop the conventional LUR models; Second, Kriging-based PM₁₀ and O₃ estimations were further added into the explanatory variables pools for building the kriging-based hybrid models; In addition, predictors variables selected by the conventional LUR and hybrid models were integrated with GWR and GTWR models to predict PM₁₀ and O₃ exposure as well, this earned a comparison of six models. In the final step, all models were used to illustrate the spatial variability of PM₁₀ and O₃.

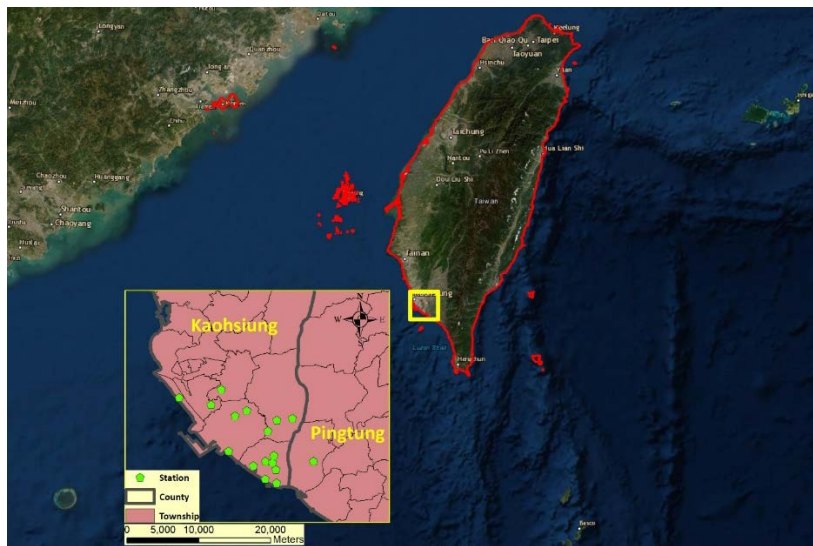


Fig. 1. Study area

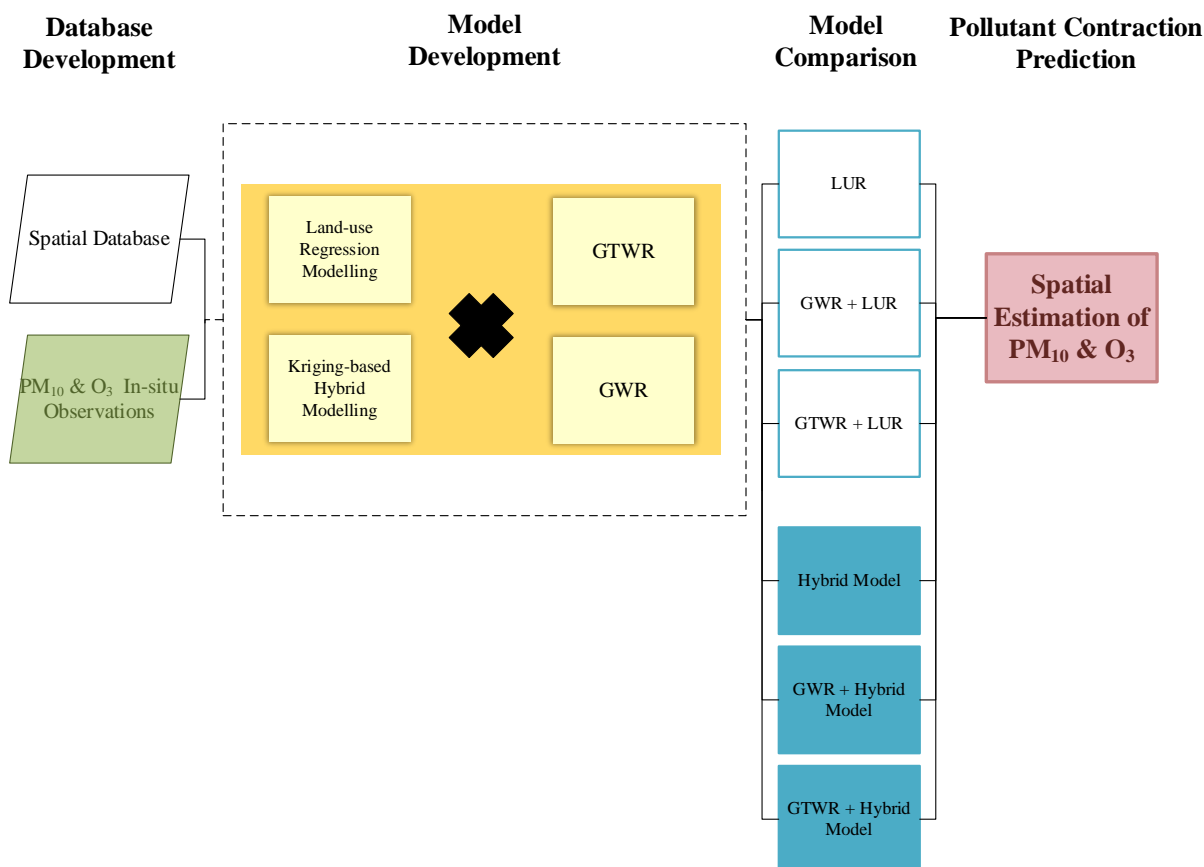


Fig. 2. Study flowchart

RESULTS

The results (Table 3) showed that, integrating the GTWR and predictor variables selected by kriging-based hybrid models have the greatest performances compared to LUR, GWR, and the combination of both with kriging-based hybrid model. R^2 obtained from the GTWR coupled with kriging-based hybrid models for PM₁₀ and O₃ was 0.96 and 0.92, respectively. Of all variables used, wind speed, pure residential area, manufacturing, park; rain-ed crop, orchard; and forest land were important predictors for PM₁₀. Whereas, wind direction, industrial area, dry crop, and orchard were variables selected to predict O₃ (Table 2; Table 3). Then, estimating PM₁₀ and O₃ concentration surfaces predicted by these 6 models. They all showed that located in the crowded and bustling place experienced higher PM₁₀ levels, and O₃ concentration is higher near farmland (Fig. 3; Fig. 4).

Table 3. All model results for PM₁₀

Variable	LUR Model	LUR Model + GWR	LUR Model + GTWR
	Coefficient	Coefficient (Median)	Coefficient (Median)
(Intercept)	+16.19***	-0.51	-2.72

NO ₂	+3.13 ^{***}	3.34	+3.35
Rainfall	-0.39 ^{***}	-0.47	-0.47
Wind speed	-8.03 ^{***}	-2.14	-0.94
Fall	+3.22 ^{***}	+4.44	+4.28
Pure residential area 500m	+9.08×10 ^{-3**}	+0.02	+0.01
Manufacturing 1250m	+5.99×10 ^{-3***}	+7.71×10 ⁻³	+7.17×10 ⁻³
Funeral facility nearest distance	-7.29×10 ⁻⁴	-7.46×10 ⁻⁴	-9.19×10 ⁻⁴
Park _{250m}	-0.02 ^{**}	-0.03	-7.03×10 ⁻³
Park _{2500m}	-0.20 ^{***}	-0.14	-0.16
Rain-ed crop _{150m}	+7.53×10 ^{-3*}	+7.59×10 ⁻³	+0.01
Orchard _{1250m}	+0.02 ^{**}	+0.02	+0.02
All farmland _{50m}	+6.84×10 ^{-3**}	+3.88×10 ⁻³	+5.80×10 ⁻³
Forest land _{4000m}	-0.04 ^{***}	-0.03	-0.03

Model Performance	R² = 0.89 ADJ R² = 0.89 RMSE = 7.29	R² = 0.9 ADJ R² = 0.9	R² = 0.91 ADJ R² = 0.9
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Variable	Hybrid Model	Hybrid Model + GWR	Hybrid Model + GTWR
	Coefficient	Coefficient (Median)	Coefficient (Median)
(Intercept)	+1.55	+5.69	-6.23
PM₁₀ (Kriging-based)	+0.97 ^{***}	+0.99	+0.99
Wind speed	-3.75 ^{***}	-0.93	-0.22
Pure residential area 500m	+0.02 ^{***}	+0.02	+0.01
Manufacturing 1250m	+7.53×10 ^{-3***}	+8.30×10 ⁻³	+7.42×10 ⁻³
Park 2500m	-0.03 [*]	-0.02	-0.06
Rain-ed crop 150m	+0.01 ^{***}	+0.01	+0.01
Orchard 1250m	+0.03 ^{***}	+0.03	+0.04
Forest land 4000m	-0.02 ^{***}	-0.02	-0.03
Model Performance	R² = 0.95 ADJ R² = 0.95 RMSE = 5.2	R² = 0.94 ADJ R² = 0.94	R² = 0.96 ADJ R² = 0.96

* $p < 0.05$; ** $p < 0.01$; *** $p < 0.001$

Table 3. All model results for O₃

Variable	LUR Model	GWR	GTWR
	Coefficient	Coefficient (Median)	Coefficient (Median)
(Intercept)	-73.71	-148.51	-2.34
Relative humidity	-0.45 ^{***}	-0.34	-0.49
Atmospheric pressure	+0.13 [*]	+0.20	+0.05
Wind direction	-0.02 ^{***}	-0.02	-0.02
Fall	+8.23 ^{***}	+8.76	+7.56
Industrial area _{nearest} distance	+1.11×10 ^{-3***}	+1.03×10 ⁻³	+7.86×10 ⁻⁴
Dry crop _{150m}	+0.01 ^{***}	+0.01	+0.01
Orchard _{50m}	+7.22×10 ^{-3**}	+7.61×10 ⁻³	.
Orchard _{1000m}	-0.03 ^{***}	-0.03	-0.02
NDVI _{mean5000m}	+8.74×10 ^{-4**}	+1.10×10 ⁻³	+1.27×10 ⁻³
Model Performance	R² = 0.42 ADJ R² = 0.41 RMSE = 5.11	R² = 0.44 ADJ R² = 0.42	R² = 0.57 ADJ R² = 0.57
Variable	Hybrid Model	GWR	GTWR
	Coefficient	Coefficient (Median)	Coefficient (Median)
(Intercept)	+1.39 ^{**}	+0.68	-0.07
O₃ (Kriging-based)	+0.98 ^{***}	+0.99	+0.99
Wind direction	-0.01 ^{***}	-8.65×10 ⁻³	-5.39×10 ⁻³
Industrial area _{nearest} distance	+0.24×10 ^{-3***}	+1.29×10 ⁻³	+1.26×10 ⁻³
Dry crop _{150m}	+8.50×10 ^{-3***}	+9.69×10 ⁻³	+8.19×10 ⁻³
Orchard _{50m}	+8.64×10 ^{-3***}	+9.18×10 ⁻³	.
Orchard _{1000m}	-0.03 ^{***}	-0.03	-0.02
Model Performance	R² = 0.87 ADJ R² = 0.87 RMSE = 2.4	R² = 0.87 ADJ R² = 0.87	R² = 0.92 ADJ R² = 0.92

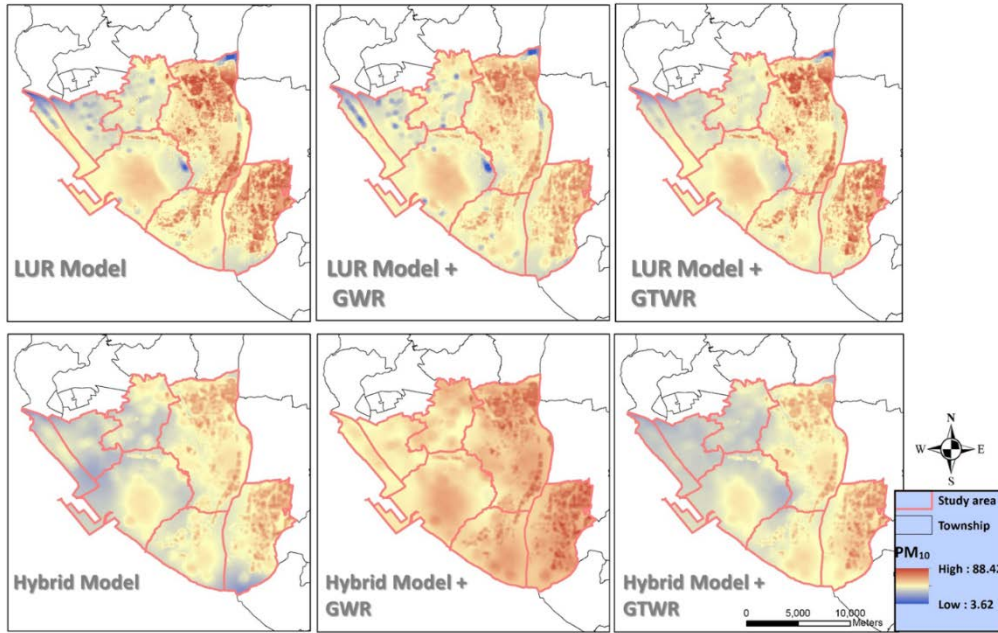


Fig. 3. Averaged PM₁₀ concentration surfaces predicted by the all models

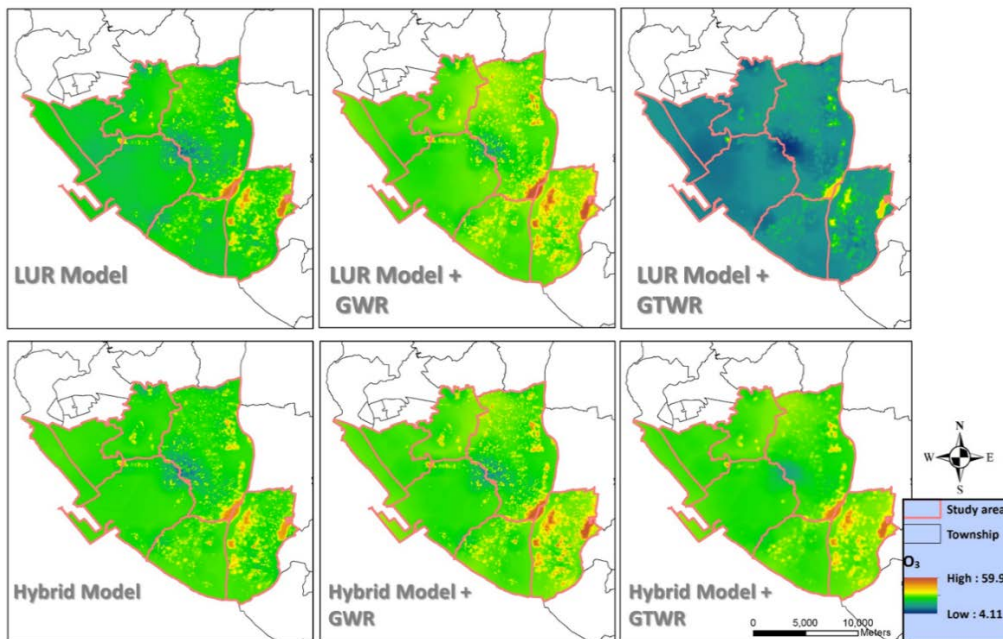


Fig. 4. Averaged O₃ concentration surfaces predicted by the all models

3. CONCLUSION

The estimated R² assured the robustness of the performance of integrated GTWR and kriging-based hybrid models on predicting temporal-spatial variability of PM₁₀ and O₃ in this study.

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REFERENCE

- Aguilera, I., Eeftens, M., Meier, R., Ducret–Stich, R. E., Schindler, C., Ineichen, A., et al. 2015. Land use regression models for crustal and traffic-related PM_{2.5} constituents in four areas of the SAPALDIA study. *Environmental Research*, 140, pp. 377-384.
- Bayraktar, H., Turalioglu, F. S., 2005. A Kriging-based approach for locating a sampling site-in the assessment of air quality. *Stochastic Environmental Research and Risk Assessment*, 9, pp. 301-305.
- Brook, R. D., Rajagopalan, S., Pope, C. A., Brook, J. R., Bhatnagar, A., Diez-Roux, A. V., et al. 2010. Particulate matter air pollution and cardiovascular disease: An update to the scientific statement from the American Heart Association. *Circulation*, 121, pp. 2331-2378.
- Chua, H.-J., Kong, S.-J., Chang, C.-H., 2018. Spatio-temporal water quality mapping from satellite images using geographically and temporally weighted regression. *Int J Appl Earth Obs Geoinformation*, 65, pp. 1-11.
- Clougherty, J. E., Kheirbek, I., Eisl, H. M., Ross, Z., Pezeshki, G., Gorczynski, J. E., et al. 2013. Intra–urban spatial variability in wintertime street-level concentrations of multiple combustion-related air pollutants: The New York City Community Air Survey (NYCCAS). *Journal of Exposure Science and Environmental Epidemiology*, 23(3), pp. 232-240.
- Cui, L., Li, R., Zhang, Y., Meng, Y., Zhao, Y., Fu, H., 2019. A geographically and temporally weighted regression model for assessing intra-urban variability of volatile organic compounds (VOCs) in Yangpu district, Shanghai. *Atmospheric Environment*, 213, pp. 746-756.
- Guo, Y., Tang, Q., Gong, D.-Y., Zhang Z., 2017. Estimating ground-level PM_{2.5} concentrations in Beijing using a satellite-based geographically and temporally weighted regression model. *Remote Sensing of Environment*, 198, pp. 140-149.
- Hoek, G., Beelen, R., Hoogh, K. D., Vienneau, D., Gulliver, J., Fischer, P., et al. 2008. A review of land–use regression models to assess spatial variation of outdoor air pollution. *Atmospheric Environment*, 42, pp. 7561-7578.
- Jerrett, M., Burnett, R. T., Ma, R., Pope, C. A., Krewski, D., Newbold, K. B., et al. 2005. Spatial Analysis of Air Pollution and Mortality in Los Angeles. *Epidemiology* 16(6), pp. 727-736.
- Maa, X., Zhang, J., Ding, C., Wang, Y., 2018. A geographically and temporally weighted regression model to explore the spatiotemporal influence of built environment on transit ridership. *Computers, Environment and Urban Systems*, 70, pp. 113-124.
- Michanowicz, D. R., Shmool, J. L., Cambal, L., Tunno, B. J., Gillooly, S., Hunt, M. J. O., et al. 2016. A hybrid land use regression/line–source dispersion model for predicting intra–urban NO₂. *Transportation Research Part D: Transport and Environment* 43, pp. 181-191.

- Pope, C. A., Burnett, R. T., Thu, M. J., Calle, E. E., Krewski, D., Ito, K., et al. 2002. Lung cancer, cardiopulmonary mortality, and long-term exposure to fine particulate air pollution. *JAMA The Journal of the American Medical Association*, 287, pp.1132-1141.
- Pope, C. A., Dockery, D. W., 2006. Health effects of fine particulate air pollution: lines that connect. *Journal of the Air & Waste Management Association*, 56(6), pp. 709-742.
- Sabaliauskas, K., Jeong, C. H., Yao, X., Reali, C., Sun, T., Evans, G. J., 2015. Development of a land-use regression model for ultrafine particles in Toronto, Canada. *Atmospheric Environment*, 110, pp. 84-92.
- Shi, L., Zanobetti, A., Kloog, I., Coull, B. A., Koutrakis, P., Melly, S. J., et al. 2016. Low-concentration PM_{2.5} and mortality: estimating acute and chronic effects in a population-based study. *Environmental Health Perspectives*, 124(1), pp. 46-52.
- Su, J. G., Jerrett, M., Nazelle, A. D., Wolch, J., 2011. Does exposure to air pollution in urban parks have socioeconomic, racial or ethnic gradients? *Environmental Research* 111, pp. 319-328.
- Tunno, B. J., Michanowicz, D. R., Shmool, J. L., Kinnee, E., Cambal, L., Tripathy, S., et al. 2016. Spatial variation in inversion-focused vs 24-h integrated samples of PM_{2.5} and black carbon across Pittsburgh, PA. *Journal of Exposure Science and Environmental Epidemiology*, 26(4), pp. 365-376.
- Wang, R., Henderson, S. B., Sbihi, H., Allen, R. W., Brauer, M., 2013. Temporal stability of land use regression models for traffic-related air pollution. *Atmospheric Environment*, 64, pp. 312-319.
- Wu, C. D., Chen, Y. C., Pan, W. C., Zeng, Y. T., Chen, M. J., Guo, Y. L., et al. 2017. Land-use regression with long-term satellite-based greenness index and culture-specific sources to model PM_{2.5} spatial-temporal variability. *Environmental Pollution*, 224, pp. 148-157.